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Key Points:

- Ammonium oxidation is the dominant N_2O production pathway in suboxic Saanich Inlet
- Addition of nitrate and nitrite stimulates N_2O production in anoxic, N_2O -depleted deep water
- Oxygenation of anoxic water enhances N_2O production via ammonium oxidation and increases water column N_2O supersaturation level

Supporting Information:

- Supporting Information S1

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Temporal and Vertical Oxygen Gradients Modulate Nitrous Oxide Production in a Seasonally Anoxic Fjord: Saanich Inlet, British Columbia

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Abstract Nitrous oxide (N_2O) is a strong greenhouse gas and an ozone depleting agent. In marine environments, N_2O is produced biologically via ammonium oxidation, nitrite, and nitrate reduction. The relative importance of these principle production pathways is strongly influenced by oxygen availability. We conducted ^{15}N tracer experiments of N_2O production in parallel with measurements of N_2O concentration and natural abundance isotopes/isotopomers in Saanich Inlet, a seasonally anoxic fjord, to investigate how temporal and vertical oxygen gradients regulate N_2O production pathways and rates. In April, June, and August 2018, the depth of the oxic-anoxic interface (dissolved oxygen = $2.5 \mu\text{mol L}^{-1}$ isoline) progressively deepened from 110 to 160 m. Within the oxygenated and suboxic water column, N_2O supersaturation coincided with peak ammonium oxidation activity. Conditions in the anoxic deep water were potentially favorable to N_2O production from nitrate and nitrite reduction, but N_2O undersaturation was observed indicating that N_2O consumption exceeded rates of production. In October, tidal mixing introduced oxygenated water from outside the inlet, displacing the suboxic and anoxic deep water. This oxygenation event stimulated N_2O production from ammonium oxidation and increased water column N_2O supersaturation while inhibiting nitrate and nitrite reduction to N_2O . Results from ^{15}N tracer incubation experiments and natural abundance isotopomer measurements both implicated ammonium oxidation as the dominant N_2O production pathway in Saanich Inlet, fueled by high ammonium fluxes ($0.6\text{--}3.5 \text{ nmol m}^{-2} \text{ s}^{-1}$) from the anoxic depths. Partial denitrification contributed little to water column N_2O production because of low availability of nitrate and nitrite.

1. Introduction

Nitrous oxide (N_2O) is an important atmospheric trace gas that regulates Earth's climate through catalysis of stratospheric ozone depletion (Crutzen, 1970) and absorption of long wave radiation (Yung et al., 1976). Globally, N_2O is mostly produced by biological activities (Codispoti, 2010), such that understanding the N_2O cycling pathways and associated regulating factors is a prerequisite for N_2O budget estimates, particularly in the marine environment (Bange et al., 2019; Wilson et al., 2018). The biological production of N_2O is attributed to nitrogen cycling processes, during which N_2O can be produced as a by-product of aerobic oxidation of ammonium (NH_4^+) to nitrite (i.e., the first step of nitrification) and as an important intermediate during the anaerobic reduction of nitrate (NO_3^-) and nitrite (NO_2^-). Consumption of N_2O occurs under anoxic conditions where heterotrophic bacteria reduce N_2O to N_2 . Over 40% of natural N_2O emissions occur in marine and freshwater environments (Ciais et al., 2013). Notably, high N_2O emissions are often associated with steep vertical oxygen gradients (oxycines), such as those found in open ocean oxygen minimum zones (OMZs) with nanomolar oxygen at intermediate depths, and eutrophic coastal waters (Bange et al., 1996; Buitenhuis et al., 2018). While open ocean N_2O emission estimates are relatively well constrained (Buitenhuis et al., 2018), coastal emissions are highly uncertain (Maavara et al., 2019), primarily due to highly variable spatial and temporal gradients of dissolved oxygen concentration (DO) and organic matter export (de Bie et al., 2002; Laperriere et al., 2019). Projecting future N_2O emission scenarios requires a better understanding of the spatial and temporal dynamics of N_2O production in relation to environmental variability, especially in productive coastal environments.

Saanich Inlet, a well-studied seasonally anoxic fjord, was chosen as the experimental site to investigate how temporal and vertical variability of dissolved oxygen regulates N_2O production in coastal waters. Located on southern Vancouver Island in British Columbia, Canada, Saanich Inlet has a central deep basin (~215 m) and a shallow sill (~70 m) at its north facing mouth, which restricts deepwater movement in and out of the fjord (Herlinveaux, 1962). Spring-neap tidal mixing at the seaward end of the fjord introduces new nutrients (primarily NO_3^-) to the surface waters of Saanich Inlet on a fortnightly basis (Gargett et al., 2003), resulting in high new primary production (Grundle & Juniper, 2011; Grundle et al., 2009). Rapid sinking of particulate organic matter and subsequent remineralization cause water column deoxygenation (Timothy & Soon, 2001), creating a steep oxycline on top of the anoxic bottom water where hydrogen sulfide (H_2S) accumulates (Cohen, 1978; Torres-Beltrán et al., 2017). Deepwater oxygen renewal events occur primarily in later summer and fall, when dense oxygenated water accumulates outside of the fjord and periodically flows over the sill and into the basin (Anderson & Devol, 1973; Manning et al., 2010). These features make Saanich Inlet an ideal natural laboratory to examine the effects of dissolved oxygen availability on nitrogen cycling (Bourbonnais et al., 2013; Grundle & Juniper, 2011; Ward & Kilpatrick, 1990), greenhouse gas fluxes (Capelle et al., 2018, 2019), and microbial community dynamics (Torres-Beltrán et al., 2016; Zaikova et al., 2010). The detailed biogeochemical characterizations of Saanich Inlet provide a foundation for further exploration of N_2O production pathways and the regulating environmental factors, which have not been reported. In this study, we have built upon earlier nitrogen cycling studies by conducting ^{15}N tracer incubation experiments to directly measure N_2O production via multiple pathways along vertical and temporal oxygen gradients. The analyses of natural abundance bulk N_2O isotopes (^{15}N versus ^{14}N and ^{18}O versus ^{16}O) and isotopomers (the intramolecular configuration of nitrogen isotopes within the linear N_2O molecule) provide a further indication of how the different metabolic pathways have contributed to total N_2O production in Saanich Inlet. For example, ^{15}N tracer experiments determine potential and instantaneous N_2O production rates and pathways (Ji et al., 2018), whereas N_2O isotopomeric measurements reveal the dominant N_2O production pathway integrated over the history of a water mass (Fujii et al., 2013). The isotope measurements reported here were conducted on a recently developed laser-based analytical system that is able to deliver accuracy and precision comparable to traditional isotope ratio mass spectrometer techniques (Ji & Grundle, 2019). This work combines ^{15}N tracer incubation, N_2O concentration, and natural abundance isotopic/isotopomeric measurements to (1) elucidate the N_2O dynamics in a model coastal marine system with temporal and vertical oxygen gradients, (2) examine oxygen regulation of N_2O production pathways during bottle incubations and under in situ conditions, and (3) identify the major N_2O production pathways potentially contributing to atmospheric emissions.

2. Materials and Methods

2.1. Field Sampling

Four cruises were conducted on the *MSV John Strickland* in 2018 to conduct sampling in Saanich Inlet. Our sampling site ($48^\circ38.58'\text{N}$, $123^\circ29.93'\text{W}$, depth 206 m, Figure S1) was ~10 km south of the inlet's mouth and ~5 km north of the monthly N_2O time series station reported by Capelle et al. (2018). Bimonthly sampling was conducted on 5 April, 14 June, 2 August, and 25 October 2018 to represent the transition from stable water column stratification and bottom water anoxia to deepwater oxygen renewal.

Six discrete sampling depths below the euphotic zone (75, 90, 100, 110, 130, and 160 m) were chosen to bracket the oxycline and were maintained consistently throughout the four cruises. Water samples were collected from 5-L Niskin bottles on a rosette sampler equipped with a conductivity-temperature-depth (CTD) package (SBE 19, Sea-Bird Electronics, Bellevue, WA). The CTD rosette was also equipped with an oxygen sensor (SBE 43, Sea-Bird Electronics) for real time, continuous profiling of DOs. Dissolved oxygen was also measured in discrete samples, using the Carpenter-Winkler titration method (Carpenter, 1965) for the purpose of calibrating the oxygen sensor. The discrete DO measurements generally had a reproducibility of 2% of mean values, and the oxygen sensor had a detection limit of $\sim 2.5 \mu\text{mol L}^{-1}$. We operationally define the terms *anoxic* as DO ranging from 0 to $2.5 \mu\text{mol L}^{-1}$, *suboxic* as 2.5 to $15 \mu\text{mol L}^{-1}$, and *oxygenated* as greater $15 \mu\text{mol L}^{-1}$. It should be noted that at anoxic conditions ($\text{DO} = 0\text{--}2.5 \mu\text{mol L}^{-1}$), denitrification and sulfate reduction are probably the dominant microbial metabolism, yet oxygen respiration continues to operate even at nanomolar DO level (Zakem & Follows, 2017). And thus, denitrification, sulfate reduction and

oxygen respiration coexist at suboxic depths in Saanich Inlet (Bourbonnais et al., 2013; Manning et al., 2010; Torres-Beltrán et al., 2016). Seawater subsamples for nutrient measurements were filtered (0.22- μm Sterivex™ filter, EMD Millipore, Burlington, MA) and stored at -20°C in acid-washed 60-ml high-density polyethylene bottles (20,160,060, Thermo Scientific, Waltham, MA, USA).

Duplicate samples for N_2O concentration measurements were filled from 5-L Niskin bottles into the bottom of 20-ml glass serum vials (LPP.10732, LEAP PAL Parts, Raleigh, NC, United States) and then were allowed to overflow at least three times the volume before sealing the vials with butyl septa (60180744, Thermo Scientific) and aluminum rings (60180512, Thermo Scientific). The mean volume of a crimp-sealed vial was 20.3 ± 0.06 ml. Duplicate samples for N_2O isotope and isotopomer measurements were crimp sealed in 60-ml glass serum bottles (223745, Wheaton, Millville, New Jersey, United States). Both concentration and isotope/isotopomer samples were immediately preserved with 0.05 ml of saturated mercuric chloride solution (HgCl_2). Samples were stored in the dark at 20 to 23°C for less than 4 months before laboratory analyses (described in section 2.3).

2.2. Nitrogen Tracer Experiments

For N_2O production experiments, seawater was sampled from 5-L Niskin bottles into crimp-sealed 20-ml vials using the same technique as described for the N_2O concentration samples. To facilitate tracer addition, a 1-ml N_2 headspace was created in the crimp-sealed vials. During the October sampling, the four shallower depths (75, 90, 100, and 110 m) were treated with an air headspace to maintain oxygenated conditions. Final DOs in the water phase were calculated using equilibrium concentrations between the water and gas phases in the incubation vials (Garcia & Gordon, 1992) and were later confirmed by laboratory oxygen sensor measurements in a similar experimental setup (see supporting information Text S1). The water phase DO in vials with an N_2 headspace was approximately 40% of in situ concentration, whereas air headspace vials had elevated DO ranging from 187 to $210\ \mu\text{mol L}^{-1}$, in comparison to in situ DO of $43\text{--}101\ \mu\text{mol L}^{-1}$ (see supporting information Table S1 for DO of all incubation experiments). Incubation experiments with ^{15}N tracers were applied to quantify rates of N_2O production from NH_4^+ oxidation, NO_2^- , and NO_3^- reduction. Three suites of ^{15}N tracer solutions ($^{15}\text{NH}_4^+$ plus $^{14}\text{NO}_2^-$, $^{15}\text{NO}_2^-$ plus $^{14}\text{NH}_4^+$, $^{15}\text{NO}_3^-$ plus $^{14}\text{NH}_4^+$, and $^{14}\text{NO}_2^-$, Sigma-Aldrich, St. Louis, Missouri, United States) were applied to enrich $^{15}\text{NH}_4^+$, $^{15}\text{NO}_2^-$, and $^{15}\text{NO}_3^-$ to 0.5, 0.5 and $1.0\ \mu\text{mol L}^{-1}$ (final concentration, tracer ^{15}N atom % = 99%), respectively, and increased concentrations of $^{14}\text{NH}_4^+$, $^{14}\text{NO}_2^-$, or $^{14}\text{NO}_3^-$ by 0.5, 0.5, and $1.0\ \mu\text{mol L}^{-1}$, respectively. Tracers were dissolved in deionized water, and the solutions were flushed with N_2 before adding 0.1 ml into each vial. The set of incubations with $^{15}\text{NH}_4^+$ plus $^{14}\text{NO}_2^-$ was also used for measuring nitrification rates (oxidation of NH_4^+ to NO_2^- and NO_3^-), which allowed us to estimate the yield of N_2O from nitrification, defined as the molar nitrogen ratio of N_2O production to NH_4^+ oxidation (see section 3.4). Incubations lasted 12 to 18 hr in temperature-controlled chambers ($\pm 1^\circ\text{C}$ of in situ temperature), during which duplicate samples were preserved every 6 to 9 hr (3 time points in total) with 0.05 ml of saturated HgCl_2 solution. Preserved samples were stored in the dark at 20 to 23°C for less than 4 months before performing N_2O isotopic/isotopomeric analyses (see section 2.3).

2.3. Laboratory Analyses

Concentrations of dissolved inorganic nitrogen species (NH_4^+ , NO_2^- and NO_3^-) were determined as follows: NH_4^+ was measured fluorometrically by reaction with orthophthaldialdehyde (Holmes et al., 1999), with a detection limit of $0.02\ \mu\text{mol L}^{-1}$; NO_2^- was treated with the Griess reagent and measured colorimetrically (Hansen & Koroleff, 1999), with a detection limit of $0.005\ \mu\text{mol L}^{-1}$ (using a 10-cm cuvette); and $\text{NO}_3^- + \text{NO}_2^-$ was measured using the cadmium reduction method (Hansen & Koroleff, 1999), with a detection limit of $0.02\ \mu\text{mol L}^{-1}$. Although not experimentally measured, the presence of hydrogen sulfide (H_2S) was indicated by its recognizable smell. However, samples for nutrient analyses were filtered in open air, and H_2S was probably oxidized. It is therefore unlikely for H_2S to have interfered with $\text{NO}_3^- + \text{NO}_2^-$ measurements.

Analyses of N_2O concentration and isotope/isotopomer ratios were performed using a purge-and-trap module coupled to a cavity ring-down spectrometer (PT-CRDS) recently developed in the Grundle Laboratory. The PT-CRDS is an automated, laser-based system that provides comparable accuracy and precision to gas chromatography and isotope ratio mass spectrometer techniques for N_2O concentration and

isotope/isotopomer ratio measurements, respectively. A full description of this technique is outlined in Ji and Grundle (2019). Isotopic analysis determines the relative abundances of nitrogen isotopes ($\delta^{15}\text{N}_{\text{bulk}}$) and oxygen isotopes ($\delta^{18}\text{O}$) of the entire N_2O pool, whereas isotopomeric analysis discerns the intramolecular nitrogen isotope substitution (isotopomers) on the linear asymmetric N_2O molecule ($\text{N}=\text{N}=\text{O}$). By convention, the relative abundance of ^{15}N substitutions of the central ($^{14}\text{N} = ^{15}\text{N} = ^{16}\text{O}$) and terminal positions ($^{15}\text{N} = ^{14}\text{N} = ^{16}\text{O}$) are denoted as $\delta^{15}\text{N}_2\text{O}_\alpha$ and $\delta^{15}\text{N}_2\text{O}_\beta$, respectively. Briefly, the PT-CRDS analyses employed the following methodology: Dissolved N_2O in sample vials was extracted by N_2 purging and cryo-trapped using liquid nitrogen at working temperature of 130 ± 10 K (note that N_2O boiling temperature is 185 K). Heating to 480 K released the gaseous N_2O , after which concentrations and isotope/isotopomer ratios were measured by quantitative light dissipation at specific wavelengths ($^{14}\text{N} = ^{15}\text{N} = ^{16}\text{O}$, $^{15}\text{N} = ^{14}\text{N} = ^{16}\text{O}$, and $^{14}\text{N} = ^{14}\text{N} = ^{18}\text{O}$ at 2195.76195, 2198.79576, and 2195.95102 cm^{-1} , respectively). Raw isotopic measurements were corrected for sample size-dependent isotopic deviations caused by varying N_2O concentrations (Ji & Grundle, 2019) and calibrated with isotopic reference materials (see Table S2). Because the optimal quantity of N_2O for highly accurate isotopic/isotopomeric analyses by PT-CRDS is 0.5–0.8 nmol- N_2O , samples from 130 and 160 m in April and 160 m samples in June did not contain enough N_2O , and, as such, those results are not reported.

The $^{15}\text{NH}_4^+$ incubation set was used for quantifying nitrification rates after measuring $\delta^{15}\text{N}-\text{N}_2\text{O}$. Nitrification rates were determined as the transfer of nitrogen from NH_4^+ to $\text{NO}_2^- + \text{NO}_3^-$ (Laperriere et al., 2019). The $\delta^{15}\text{N}$ signature of $\text{NO}_2^- + \text{NO}_3^-$ ($\delta^{15}_{\text{NO}_x^-}$) was measured using the same PT-CRDS technique outlined above, but only after first converting the $\text{NO}_2^- + \text{NO}_3^-$ to N_2O in crimp-sealed 20-ml vials using the denitrifier method (Weigand et al., 2016). A set of 20-ml vials (LPP.10732, LEAP PAL Parts, Raleigh, NC, United States) containing live bacterial concentrate *Pseudomonas chlororaphis* (ATCC® 43928™, Manassas, Virginia) in buffer solution (pH = ~7.3) were purged with nitrogen gas to create anaerobic conditions in the vials. The water samples were treated with 0.05 ml of 10 mol L^{-1} sodium hydroxide solution in order to precipitate mercury; the supernatant was transferred to the bacterial vials for complete conversion of $\text{NO}_2^- + \text{NO}_3^-$ to N_2O within 3 hr of incubation at 22 °C. As the $\text{NO}_2^- + \text{NO}_3^-$ concentrations of water samples ranged from 0.6 to 32 $\mu\text{mol L}^{-1}$, the volume of water sample injected for bacterial conversion was adjusted to accommodate 0.5 nmol- N_2O in order to achieve accurate $\delta^{15}\text{N}$ measurements by the PT-CRDS. To calibrate the measurements, isotopic reference vials were prepared by bacterial conversion of potassium nitrate USGS-35 ($\delta^{15}\text{N} = 2.7\text{‰}$ versus Air N_2) and USGS-32 ($\delta^{15}\text{N} = 180\text{‰}$ versus Air N_2).

2.4. Data Processing

Water column N_2O saturation was quantified by the N_2O excess ($\Delta\text{N}_2\text{O}$), defined as the concentration difference between measured and equilibrium values with respect to the atmosphere.

$$\Delta\text{N}_2\text{O} = [\text{N}_2\text{O}]_{\text{measured}} - [\text{N}_2\text{O}]_{\text{equilibrium}} \quad (1)$$

Equilibrium N_2O concentrations are temperature and salinity dependent and were calculated according to Weiss and Price (1980) using atmospheric N_2O concentration of 330 ppb.

The δ notation is used to quantify the relative abundance of N_2O isotopomers (Coplen, 2011) in Equations 2 and 3.

$$\delta^{15}\text{N}_2\text{O}_\alpha = \frac{(^{14}\text{N}^{15}\text{N}^{16}\text{O}/^{14}\text{N}^{14}\text{N}^{16}\text{O})}{R_{\text{ref}}} - 1, \quad R_{\text{ref}} = \frac{^{15}\text{N}}{^{14}\text{N}} \text{ of atmospheric } \text{N}_2, 0.003667 \quad (2)$$

$$\delta^{15}\text{N}_2\text{O}_\beta = \frac{(^{15}\text{N}^{14}\text{N}^{16}\text{O}/^{14}\text{N}^{14}\text{N}^{16}\text{O})}{R_{\text{ref}}} - 1, \quad R_{\text{ref}} = \frac{^{15}\text{N}}{^{14}\text{N}} \text{ of atmospheric } \text{N}_2, 0.003667 \quad (3)$$

The bulk isotopes of N_2O are defined in Equations 4 and 5.

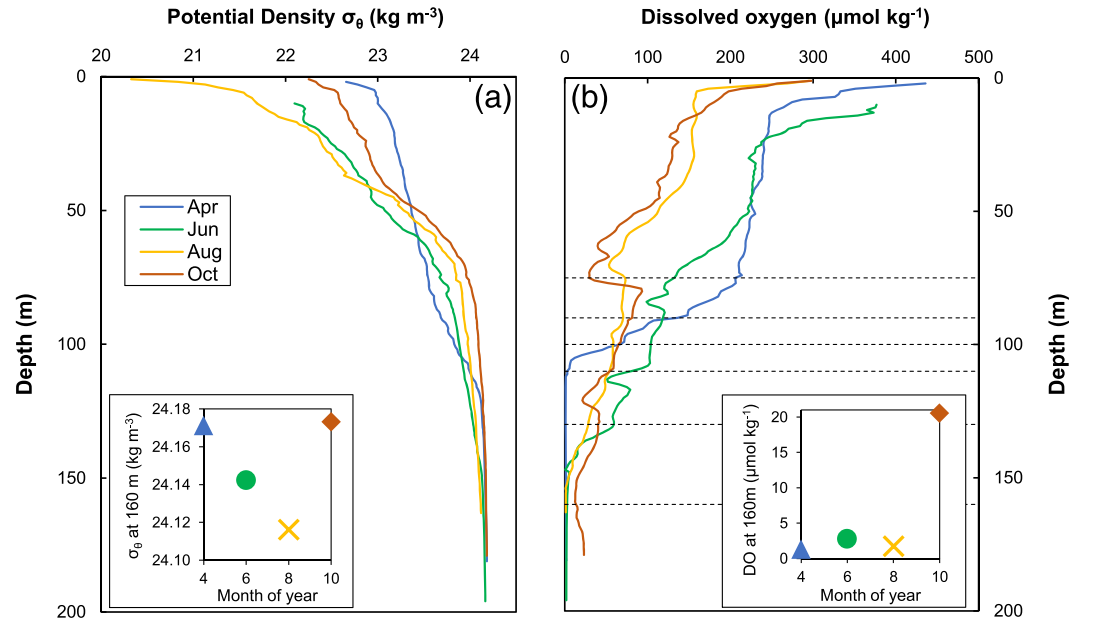


Figure 1. Profiles of water column (a) potential density and (b) dissolved oxygen at the sampling station in Saanich Inlet ($48^\circ 47.707' \text{N}$, $123^\circ 29.927' \text{W}$) during four campaigns in 2018 (April, June, August, and October). Inset figures of (a) and (b) show the potential density and dissolved oxygen at 160 m. Dashed lines in (b) indicate the sampling depths (75, 90, 100, 110, 130, and 160 m) for inorganic nitrogen analyses and N_2O production experiments.

$$\delta^{18}\text{O}-\text{N}_2\text{O} = \frac{(^{18}\text{O}/^{16}\text{O})}{R_{\text{ref}}} - 1, \quad R_{\text{ref}} = \frac{^{18}\text{O}}{^{16}\text{O}} \text{ of VSMOW, } 0.002005 \quad (4)$$

$$\delta^{15}\text{N}_{\text{bulk}}-\text{N}_2\text{O} = \frac{\delta^{15}\text{N}_\alpha + \delta^{15}\text{N}_\beta}{2} \quad (5)$$

Site preference (SP) quantifies the relative abundance of the two N_2O isotopomers according to Equation 6.

$$\text{SP} = \delta^{15}\text{N}_\alpha - \delta^{15}\text{N}_\beta \quad (6)$$

The N_2O production rate $R_{\text{N}_2\text{O}}$ ($\text{nmol-N L}^{-1} \text{ day}^{-1}$) is calculated according to Equation 7

$$R_{\text{N}_2\text{O}} = 2 \times \frac{d(^{15}\text{N}_2\text{O})/dt}{F} \quad (7)$$

where $^{15}\text{N}_2\text{O}$ represents concentration of ^{15}N -labeled N_2O molecules, $d(^{15}\text{N}_2\text{O})/dt$ represents the slope of the linear regression of $^{15}\text{N}_2\text{O}$ against time, and F represents ^{15}N enrichment of substrate (2–99% at different depths). The conversion of $\delta^{15}\text{N}_{\text{bulk}}-\text{N}_2\text{O}$ to $^{15}\text{N}_2\text{O}$ is calculated according to Equation 8

$$^{15}\text{N}_2\text{O} = [\text{N}_2\text{O}] \times \frac{(\delta^{15}\text{N}_{\text{bulk}}-\text{N}_2\text{O} + 1) \times R_{\text{ref}}}{1 + (\delta^{15}\text{N}_{\text{bulk}}-\text{N}_2\text{O} + 1) \times R_{\text{ref}}} \quad (8)$$

where $[\text{N}_2\text{O}]$ is the concentration of N_2O during the incubation experiments and $R_{\text{ref}} = 0.003667$. The detection limit of rate measurement was $0.002 \text{ nmol-N L}^{-1} \text{ day}^{-1}$.

Calculation of nitrification rate R_{nit} is according to Laperriere et al. (2019)

$$R_{\text{nit}} = \frac{d(^{15}\text{NO}_x^-)/dt}{F_{\text{NH}_4^+}^{15}} \quad (9)$$

where $d(^{15}\text{NO}_x^-)/dt$ represents the slope of linear regression of ^{15}N atom % of $\text{NO}_3^- + \text{NO}_2^-$ (abbreviated as

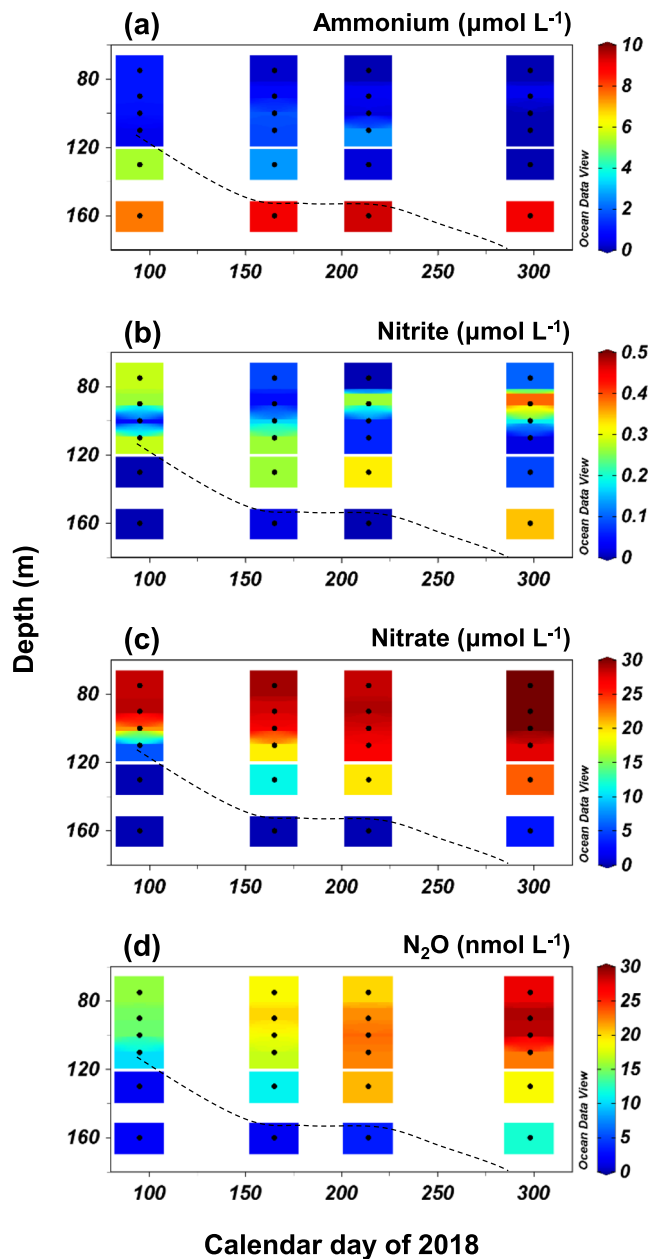


Figure 2. Concentration profiles of (a) ammonium, (b) nitrite, (c) nitrate, and (d) nitrous oxide at the sampling station (48°47.707'N, 123°29.927'W) during four sampling campaigns. Dashed line marks the approximate position of the oxic-anoxic interface (dissolved oxygen concentration = 2.5 $\mu\text{mol L}^{-1}$ isoline).

tiated the anoxic conditions at these depths. In October, the entire water column was oxygenated ($\text{DO} > 20 \mu\text{mol L}^{-1}$, Figure 1b), indicating the progressive deepening of oxic-anoxic interface ($\text{DO} = 2.5 \mu\text{mol L}^{-1}$ isoline) from April to October. Based on data from Ocean Networks Canada sensor platforms in Saanich Inlet (data not shown), oxygen renewal events appeared to have occurred in late August and mid-October. During these renewal events, dense and oxygenated water replaced the deep water inside the basin, increasing density and DO at 160 m (Figures 1a and 1b).

Inorganic nitrogen species (NH_4^+ , NO_2^- , and NO_3^-) distributed differently along the oxycline. In shallower oxygenated waters, NH_4^+ concentrations were generally $\sim 1.0 \mu\text{mol L}^{-1}$. Higher NH_4^+ concentrations (5–

NO_x^-) on time and $F_{\text{NH}_4^+}^{15\text{N}}$ represents ^{15}N fraction labeled of substrate ammonium (5–99%). The value of $^{15}\text{NO}_x^-$ is calculated according to Equation 10

$$^{15}\text{NO}_x^- = [\text{NO}_x^-] \times \frac{\left(\frac{\delta_{\text{NO}_x^-}^{15}}{1000} + 1 \right) \times R_{\text{ref}}}{1 + \left(\frac{\delta_{\text{NO}_x^-}^{15}}{1000} + 1 \right) \times R_{\text{ref}}} \quad (10)$$

where $[\text{NO}_x^-]$ is the concentration of $\text{NO}_3^- + \text{NO}_2^-$ during the incubation experiments and $\delta_{\text{NO}_x^-}^{15}$ is the $\delta^{15}\text{N}$ signature of NO_x^- measured by denitrifier method.

The nitrification N_2O yield is defined as the molar nitrogen ratio of N_2O versus nitrification. Two independent estimates of the N_2O yield are presented. First is the *rate ratio yield*, which is the ratio of directly measured rates of N_2O production versus nitrification during $^{15}\text{NH}_4^+$ incubations, and it reveals instantaneous yield from bottle incubations at a range of oxygen levels. Note that oxygen concentrations in the incubation vials were different from those in situ levels (see section 2.2). Second is the *regression yield* using the nitrate- $\Delta\text{N}_2\text{O}$ relationship, which is based on by-production of N_2O during NH_4^+ oxidation to NO_3^- in oxygenated waters below the euphotic zone, assuming stoichiometric conversion (i.e., 1 mol NH_4^+ :1 mol NO_3^-) and negligible accumulation of NO_2^- (Grundle et al., 2012; Rees et al., 2011). Such an indirect estimate reveals temporal- and spatial-averaged N_2O yield inside Saanich Inlet. To accurately characterize nitrate- $\Delta\text{N}_2\text{O}$ relationship during remineralization inside Saanich Inlet, we used NO_3^- and $\Delta\text{N}_2\text{O}$ measurements from this study and a 9-year time series data set at 40–150 m depth from a nearby station in Saanich Inlet (Torres-Beltrán et al., 2017).

3. Results

3.1. Water Column Structure and Chemistry

The water column potential density in Saanich Inlet ranged from 20.2 to 24.2 kg m^{-3} . From April to August, a progressive decrease in density of the upper 50 m signaled increased stratification. The October water column had increased density in comparison to August. At 160 m, the density level of October was similar to that of April (Figure 1a). The DOs in the upper 10 m were high and often supersaturated (up to 140%). From April to August, in situ DO decreased with depth, with anoxic conditions persisting below 110 m in April and below 150 m in June and August (Figure 1b). The 160-m water samples from April, June, and August as well as 130 m from April exhibited recognizable H_2S odor, thus providing anecdotal evidence of its presence (H_2S was not measured here), and substantiated the anoxic conditions at these depths.

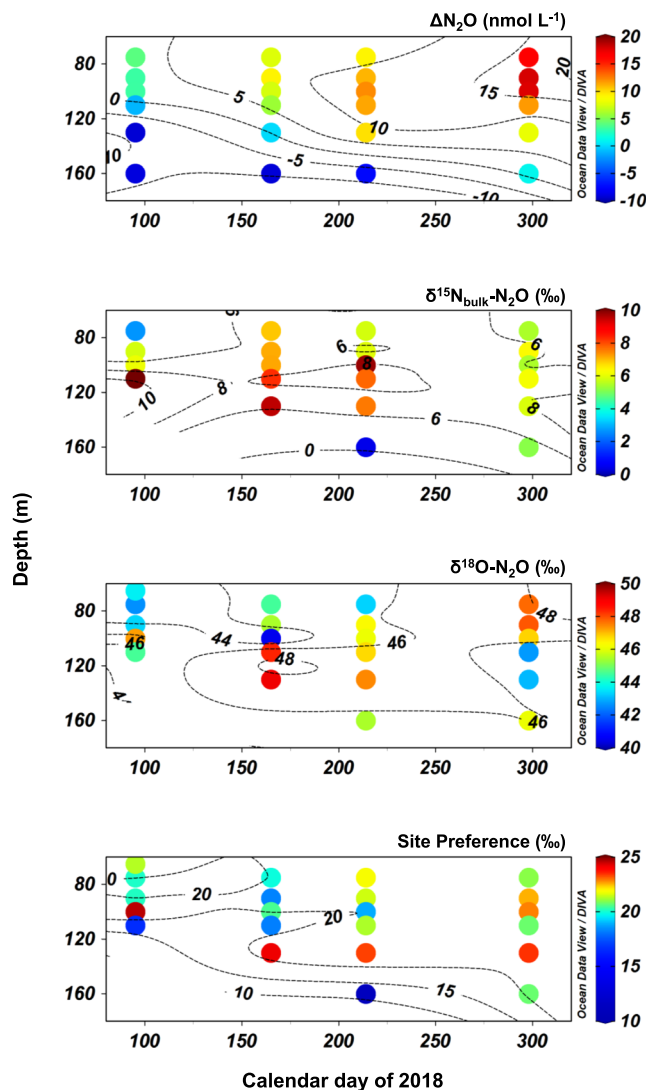


Figure 3. Water column profiles of (a) N_2O supersaturation ($\Delta\text{N}_2\text{O}$), (b) $\delta^{15}\text{N}_{\text{bulk-N}_2\text{O}}$, (c) $\delta^{18}\text{O-N}_2\text{O}$, and (d) site preference at the sampling station. Contour lines were extrapolated using DIVA algorithm from Ocean Data View software.

$\text{N L}^{-1} \text{ day}^{-1}$). No production of N_2O from NH_4^+ oxidation was detected at 130 m and below in April and at 160 m in June. In August, NH_4^+ oxidation to N_2O was detected at 160 m despite anoxic conditions (Figure 4c). In October following the renewal event, all sampling depths showed N_2O production from NH_4^+ oxidation, and the rates increased with depth. At 160 m where in situ DO was $21 \mu\text{mol L}^{-1}$, a maximum rate of N_2O production from NH_4^+ oxidation was reached at $\sim 1 \text{ nmol-N L}^{-1} \text{ day}^{-1}$ (Figure 4d), 3 to 25 times the maximum rates in April, June, and August.

Production of N_2O from NO_2^- reduction and NO_3^- reduction occurred at suboxic and anoxic conditions. Highest rates of N_2O production from NO_2^- reduction and NO_3^- reduction occurred at the oxic-anoxic interface, at 110 m in April (Figure 4e) and at 160 m in June and August (Figures 4f and 4g). These depths had pronounced N_2O undersaturation (Figure 3a) and low NO_2^- and NO_3^- concentrations that were close to the detection limit (Figures 2b and 2c). In October, low rates of N_2O production from NO_2^- and NO_3^- reduction were detected ($< 0.09 \text{ nmol-N L}^{-1} \text{ day}^{-1}$, Figure 4h); these rates generally increased with depth and with decreasing DO.

$10 \mu\text{mol L}^{-1}$) occurred in deeper anoxic waters, that is, below 130 m in April and at 160 m in June and August (Figure 2a). Nitrite concentrations were generally low ($< 0.5 \mu\text{mol L}^{-1}$) throughout the water column for all sampling dates. Local NO_2^- maximum concentrations ($\sim 0.3 \mu\text{mol L}^{-1}$) occurred above the oxic-anoxic interface (Figure 2b). The concentrations of NO_2^- were below detection at the anoxic depths. The distribution of NO_3^- showed a somewhat opposite distribution to that of NH_4^+ . In oxygenated waters between 75 and 100 m, NO_3^- concentrations were $25\text{--}30 \mu\text{mol L}^{-1}$; the concentrations decreased markedly at deeper depths, coinciding with an increase of NO_2^- above the oxic-anoxic interface (Figure 2c). Very low NO_3^- concentrations ($< 0.2 \mu\text{mol L}^{-1}$) occurred in deep anoxic waters.

3.2. Nitrous Oxide Concentration and Isotopes

Water column N_2O concentrations were $10\text{--}29 \text{ nmol L}^{-1}$ above the oxic-anoxic interface and decreased with depth. From April to October, N_2O concentrations at oxycline depths (75–110 m) progressively increased (Figure 2d). Supersaturations of N_2O ($\Delta\text{N}_2\text{O}$ up to 20 nmol L^{-1}) were observed at shallower oxycline depths from April to October (Figure 3a). Within the anoxic depths in April, June, and August, pronounced N_2O undersaturation ($\Delta\text{N}_2\text{O}$ ranged from -7.7 to -9.3 nmol L^{-1}) was observed. From August to October, N_2O concentrations at 160 m increased to near equilibrium concentration ($\Delta\text{N}_2\text{O} = 1 \text{ nmol L}^{-1}$). The oxygenated and suboxic depths had variable $\delta^{15}\text{N}_{\text{bulk-N}_2\text{O}}$ values ranging from 5‰ to 10‰ (Figure 3b) and $\delta^{18}\text{O-N}_2\text{O}$ 40‰ to 49‰ (Figure 3c). SP values were $15\text{--}25\text{‰}$ and generally increased with depth and with decreasing DO (Figure 3d). The only isotope/isotopomer sample from the anoxic depths was at 160 m in August, where the values of $\delta^{15}\text{N}_{\text{bulk-N}_2\text{O}}$, $\delta^{18}\text{O-N}_2\text{O}$, and SP were 0.4‰ , 45‰ , and 11‰ , respectively.

3.3. Nitrous Oxide Production Pathways

Production of N_2O from NH_4^+ oxidation occurred in the oxygenated and suboxic waters, coinciding with positive $\Delta\text{N}_2\text{O}$ concentrations. In April, June, and August, rates of N_2O production peaked at 100–110 m (Figures 4a–4c) where in situ DO was $50\text{--}80 \mu\text{mol L}^{-1}$. The maximum rate of N_2O production from NH_4^+ oxidation was much higher in April ($0.35 \text{ nmol-N L}^{-1} \text{ day}^{-1}$) than those in June and August ($\sim 0.05 \text{ nmol-N L}^{-1} \text{ day}^{-1}$).

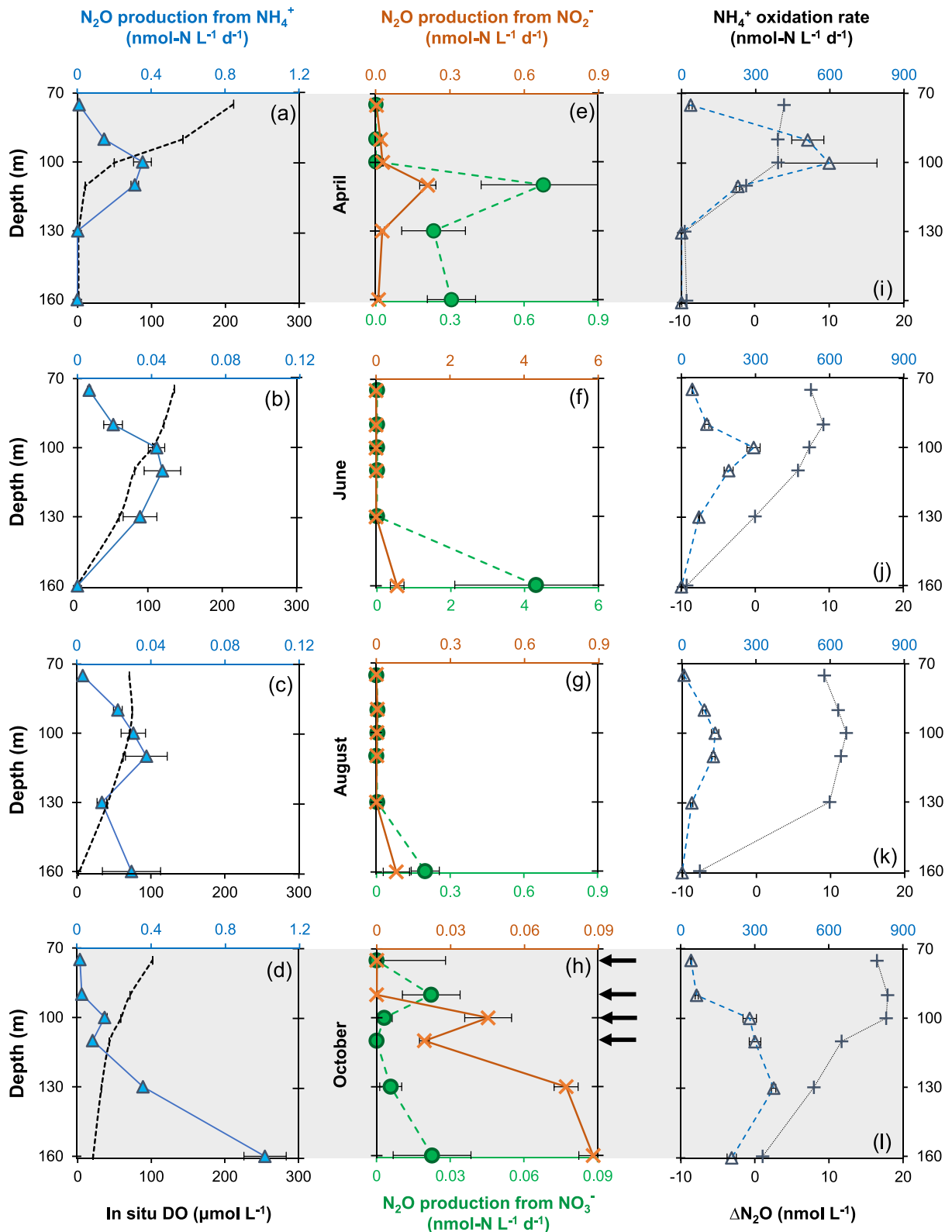


Figure 4. Profiles of N_2O production rates, nitrification rates, and corresponding dissolved oxygen concentrations. (a–d) N_2O production from NH_4^+ oxidation (triangle) and dissolved oxygen profile (dashed line). (e–h) N_2O production from NO_2^- (cross) and NO_3^- reduction (circle). (i–l) Nitrification rate (triangle) and $\Delta\text{N}_2\text{O}$ concentrations (plus sign) during four sampling campaigns (a, e, and i: April; b, f, and j: June; c, g, and k: August; d, h, and l: October). Arrows in (h) indicate that rate measurements were conducted under air headspace, thus having $\text{DO} = 180\text{--}210 \mu\text{mol L}^{-1}$, whereas the rest of the incubations were conducted under N_2 headspace, having dissolved oxygen about 40% of in situ levels (see Table S1).

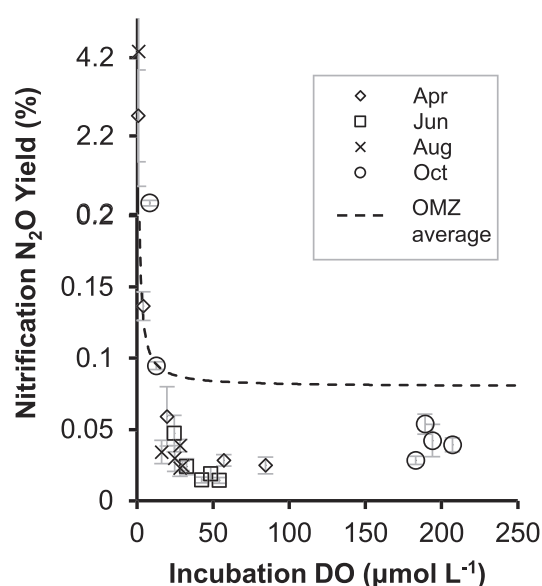


Figure 5. N_2O yield during NH_4^+ oxidation in Saanich Inlet, measured as rate ratios of N_2O production versus NH_4^+ oxidation by ^{15}N incubation experiments (referred to as *rate ratio yield*) during sampling in April (diamonds), June (squares), August (crosses), and October (circles). Dashed line represents the average yield reported by Ji et al. (2018) for the eastern tropical South and North Pacific oxygen minimum zones. Note that the scale on y axis is nonlinear.

3.4. Nitrification and Nitrous Oxide Yield

Nitrification (measured as NH_4^+ oxidation to NO_2^- plus NO_3^-) occurred in oxygenated and suboxic zones ($\text{DO} > 2.5 \mu\text{mol L}^{-1}$) from April to October. Both shallow ($<90 \text{ m}$) and deep (160 m) layers had lower rates, while the highest rates often occurred at suboxic depths close to the oxic-anoxic interface (Figures 4i–4l). Maximum rates were $130\text{--}600 \text{ nmol-N L}^{-1} \text{ day}^{-1}$ at in situ $\text{DO} = 30\text{--}140 \mu\text{mol L}^{-1}$ from April to October. Nitrification was not detected at 160 m in April and June, and rates were comparably low at 130 m in April ($0.08 \text{ nmol-N L}^{-1} \text{ day}^{-1}$, Figure 4i) and at 160 m in August ($0.68 \text{ nmol-N L}^{-1} \text{ day}^{-1}$, Figure 4l). Two independent estimates of N_2O yield during nitrification are compared as follows. The rate ratio yield ranged from 0.014 to 0.059% at $\text{DO} = 15\text{--}210 \mu\text{mol L}^{-1}$ during bottle incubations and increased significantly to $4.3 \pm 3.4\%$ at $\text{DO} < 1 \mu\text{mol L}^{-1}$ in a nonlinear fashion (Figure 5). Using nitrate concentration and $\Delta\text{N}_2\text{O}$ data in this study and from time series observations from Torres-Beltrán et al. (2017), the regression yield was $0.045 \pm 0.006\%$, which is within the range of the rate ratio yield from direct measurements at $\text{DO} > 15 \mu\text{mol L}^{-1}$ ($0.014\text{--}0.059\%$).

4. Discussion

Saanich Inlet is a model coastal system with vertical and temporal gradients of oxygen and inorganic nitrogen that can be ideal for investigating N_2O dynamics. Incubation experiments with ^{15}N -labeled substrates directly measured the rates of N_2O production via NH_4^+ oxidation,

NO_2^- reduction, and NO_3^- reduction under variable DO gradients. This allows for an objective assessment of the relative contributions of oxidative and reductive N_2O production pathways, as well as the regulatory effects of DO. Evidently, NH_4^+ oxidation was the dominant N_2O production pathway throughout our sampling period under oxygenated conditions ($\text{DO} > 15 \mu\text{mol L}^{-1}$, Figure 6a). Incubation experiments demonstrating potential N_2O production rates from the oxycline to anoxic depths are important for interpreting temporal and vertical $\Delta\text{N}_2\text{O}$ profiles. Samples with positive $\Delta\text{N}_2\text{O}$ concentrations generally co-occurred with higher rates of N_2O production via NH_4^+ oxidation than rates of NO_2^- and NO_3^- reduction (Figure 6b). Active NO_2^- and NO_3^- reduction to N_2O were mostly confined to suboxic and anoxic waters ($\text{DO} < 15 \mu\text{mol L}^{-1}$). Furthermore, N_2O was undersaturated at anoxic depths, indicating net N_2O consumption as a result of complete denitrification. Both NO_3^- and NO_2^- reduction to N_2O can be regarded as partial denitrification; however, it has been shown that the two pathways were performed independently by denitrifying cells, and intracellular exchange of NO_2^- was not detected (Ji et al., 2018). Thus, rates of NO_3^- and NO_2^- reduction to N_2O were measured separately, and it is unnecessary for the two sets of rates to match. Furthermore, these potential rates were detected at depths with near depletion of NO_2^- and NO_3^- (Figures 2b and 2c); thus, addition of ^{15}N -labeled NO_2^- and NO_3^- substrates could stimulate N_2O production via partial denitrification mediated by active denitrifying microbial communities (Capelle et al., 2018; Zaikova et al., 2010) capable of short-term response to inputs of electron acceptors.

As previously explained, ^{15}N incubation experiments were performed under altered DO, and this had minimal effect on determining the major N_2O production pathway in the water column. Potential microbial activities in the water column were demonstrated by these measured rates, which can be comparable provided that they were conducted under similar experimental conditions (e.g., comparable rates measured from N_2 -headspace vials). For example, during incubation experiments in October, the 130 and 160 m samples had decreased DO with respect to in situ values (Table S1), which may have stimulated N_2O production from NH_4^+ oxidation. The oxycline samples ($75\text{--}110 \text{ m}$) were treated with air headspace and had higher DO than in situ levels. While it is possible that N_2O production from NH_4^+ oxidation was inhibited by elevated DO (see section 3.3), these October samples from 75 to 110 m also showed higher rates ($0.02\text{--}0.15 \text{ nmol-N L}^{-1} \text{ day}^{-1}$) in comparison to August samples from the same depths, which were conducted under N_2

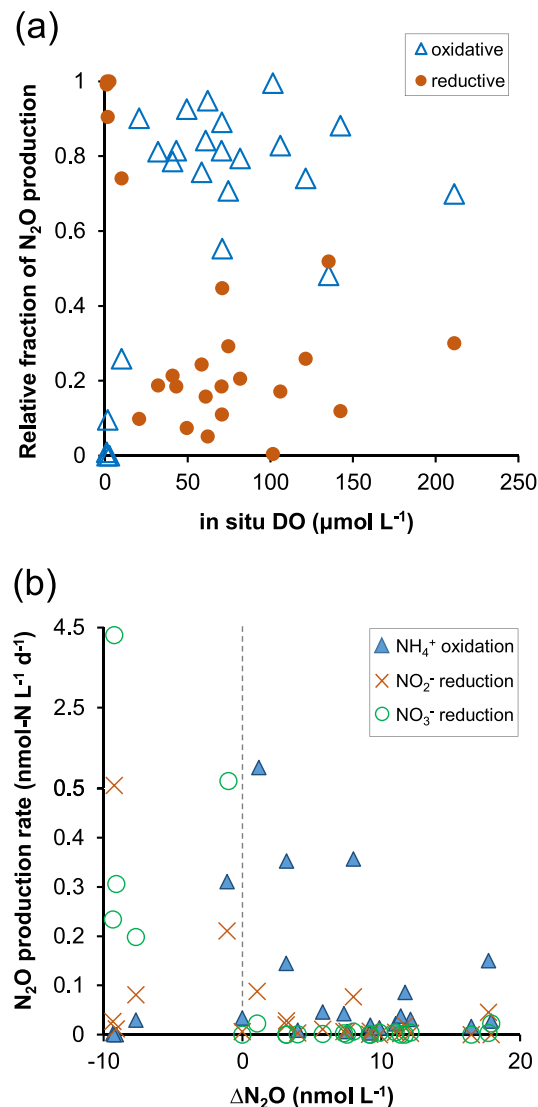


Figure 6. (a) The fraction of N₂O production from oxidative pathway (NH₄⁺ oxidation, open triangle) and reductive pathway (NO₂⁻ + NO₃⁻ reduction, filled circle) plotted against in situ dissolved oxygen concentration. (b) Rates of N₂O production via NH₄⁺ oxidation (triangle), NO₂⁻ reduction (cross), and NO₃⁻ reduction (circle) plotted against N₂O supersaturation (ΔN₂O). Dashed line marks saturated N₂O concentration with respect to atmosphere. Note that the scale on y axis is nonlinear.

headspace. Production of N₂O from NO₂⁻/NO₃⁻ reduction could be overestimated because of lowered DO during the incubation experiments with N₂ headspace. Thus, the relative importance of partial denitrification to total N₂O production could be overestimated, further substantiating our conclusion that NH₄⁺ oxidation was the dominant N₂O production pathway.

Two additional factors should be acknowledged when interpreting the rate measurements obtained from ¹⁵N tracer incubations. (1) Variable in situ NH₄⁺, NO₂⁻ and NO₃⁻ concentrations resulted in 2–99% ¹⁵N enrichment of substrates at the start of incubations. Therefore, rate measurements of N₂O production from NH₄⁺ oxidation (and nitrification) above 100 m depth and all of N₂O production from NO₂⁻ reduction and rates of NO₃⁻ reduction at 160 m (see section 3.3) should reflect microbial potential at these depths. (2) Dilution of ¹⁵N tracers during incubation experiments results in underestimation of N₂O production rates. Variable dilution of ¹⁵NO₂⁻ was estimated (0–27%) using nitrification rate measurements. Dilution effects for other sets of tracer additions (¹⁵NH₄⁺ and ¹⁵NO₃⁻) were not estimated, and rates should be considered conservative.

The SP value of N₂O is not merely an index characterizing the relative abundance of δ¹⁵N₂O_α over δ¹⁵N₂O_β; it also provides an additional, independent constraint of N₂O production pathways with respect to ¹⁵N incubation experiments. The SP value can also serve as a quantitative indicator for the fraction of N₂O produced via oxidative and reductive pathways. Microbial N₂O production via NH₄⁺ oxidation and NO₂⁻ and/or NO₃⁻ reduction has SP values of 30 ± 5‰ and 0 ± 5‰, respectively, and these values are independent of substrate isotopic composition across a wide range of environments (Toyoda et al., 2017). During N₂O production, the N₂O isotopes and SP signatures follow mass conservation. The mass balance model proposed by Fujii et al. (2013) quantified the SP signature of N₂O produced (SP_{produced}) by the linear regression of the observed isotope values (SP_{observed}) on the inverse N₂O concentration (1/[N₂O]_{measured}). Saanich Inlet samples showed an SP_{produced} value of 23 ± 0.6‰ (or 24 ± 1‰ without an outlier data point, Figure 7a), lending further support to NH₄⁺ oxidation being the predominant N₂O production pathway. Here, ¹⁵N tracer incubation experiments and natural abundance SP values permitted a parallel view of N₂O production pathways in bottle incubations in comparison to the predominant

production pathways in situ. The fraction of N₂O produced via NH₄⁺ oxidation determined by ¹⁵N tracer experiments was plotted against SP values from the corresponding samples (Figure 7b). Higher contributions from NH₄⁺ oxidation results in SP values close to 25‰, whereas higher contribution from reductive pathways results in SP values close to 10‰. These data can be explained by a mixing model with two end members: one with SP = 25 ± 5‰ representing N₂O production from NH₄⁺ oxidation and the other with SP = 10 ± 5‰, respectively representing N₂O production from NO₂⁻ and/or NO₃⁻ reduction. Additional studies are needed to test the robustness of using SP signatures to quantitatively determine the fraction of N₂O produced via oxidative and reductive pathways in Saanich Inlet and elsewhere.

In all, the two independent methods, natural abundance isotopomers and ¹⁵N tracer incubation experiments, delivered consistent results demonstrating that the dominant N₂O production pathway in Saanich Inlet is NH₄⁺ oxidation. A notable difference between Saanich Inlet and open ocean OMZs is high NH₄⁺

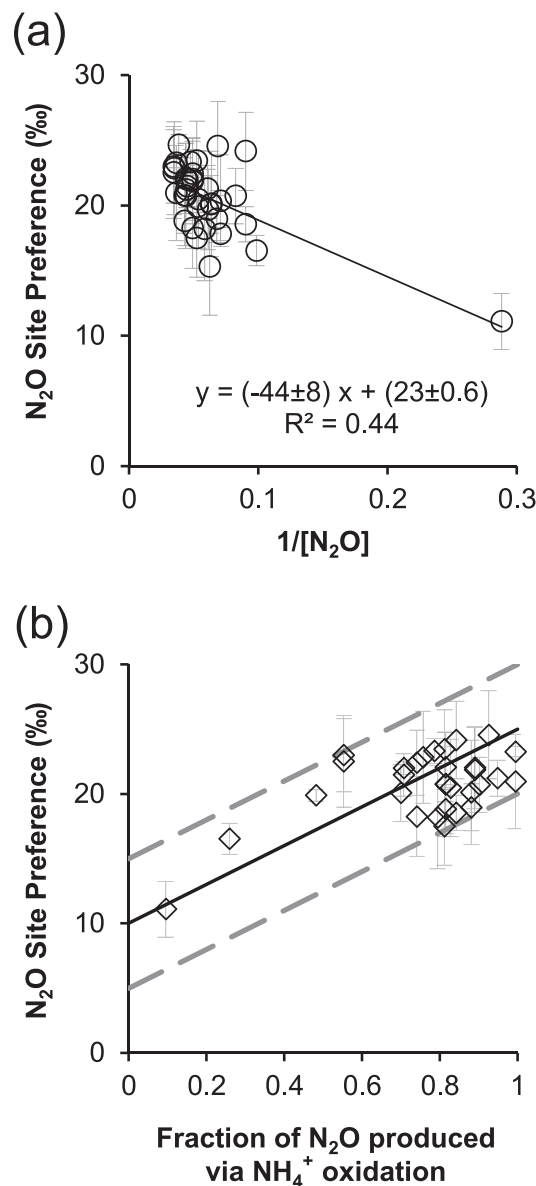


Figure 7. (a) Linear regression of N₂O site preference on inverse N₂O concentrations. Note: Without the outlier data point near the right edge of the plot, regression has a y intercept of 24 ± 1 . (b) N₂O site preference (SP) measured in situ plotted against the fraction of N₂O production from NH₄⁺ oxidation measured by ¹⁵N incubation experiments. Solid line represents mixing between two end members, one with SP signature of 25 ‰ representing N₂O production from NH₄⁺ oxidation and the other with SP signature of 10 ‰ representing N₂O production from NO₂⁻ or NO₃⁻ reduction. Two dashed lines represent upper and lower SP values of N₂O when mixing between NH₄⁺ oxidation and NO₂⁻ or NO₃⁻ reduction having $\pm 5\%$ variation in their respective SP signatures.

fluxes generated by organic matter remineralization from the anoxic depth in Saanich Inlet (Bourbonnais et al., 2013). Using NH₄⁺ concentration profiles from this study and assuming a diffusion constant of $1.0 \text{ m}^2 \text{ day}^{-1}$ previously estimated in Saanich Inlet 130–160 m depth (Louca et al., 2016), upward NH₄⁺ fluxes from the anoxic zone ranged from 0.6 to $3.5 \text{ nmol m}^{-2} \text{ s}^{-1}$. Such NH₄⁺ fluxes are similar to the Black Sea water column with sulfidic bottom water (Fuchsman et al., 2008). Similarly, NH₄⁺ oxidation was the dominant N₂O production pathway in the Black Sea (Westley et al., 2006).

Availability of oxygen is an important factor regulating N₂O yield during nitrification. Short-term DO changes (e.g., <3 hr) during incubation experiments can significantly change the yield (Ji et al., 2018; Löscher et al., 2012). In Saanich Inlet, the rate ratio yield increases with decreasing oxygen concentrations, particularly under suboxic conditions ($\text{DO} < 15 \text{ } \mu\text{mol L}^{-1}$). Saanich Inlet samples had rate ratio yield values similar to open ocean OMZ averages at $\text{DO} < 15 \text{ } \mu\text{mol L}^{-1}$; however, the yield values are generally lower than OMZ averages at $\text{DO} > 15 \text{ } \mu\text{mol L}^{-1}$ (Figure 5). To better estimate nitrification N₂O yield using nitrate- $\Delta\text{N}_2\text{O}$ relationship, we excluded NO₃⁻ and $\Delta\text{N}_2\text{O}$ data from the surface euphotic zone and deep anoxic layer and used measurements from this study and a 9-year time series data set at 40–150 m depth from a nearby station in Saanich Inlet (Torres-Beltrán et al., 2017), for two reasons: (1) Both NO₃⁻ and $\Delta\text{N}_2\text{O}$ can be transported into Saanich Inlet by renewal water, which can alter the nitrate- $\Delta\text{N}_2\text{O}$ relationship especially for samples at depths below 160 m (Anderson & Devol, 1973; Capelle et al., 2018). As such, N₂O yield estimates could be altered by signals outside the inlet. (2) Both phytoplankton activity in the euphotic zone (Grundle et al., 2009) and nitrogen loss processes in the anoxic layer (Bourbonnais et al., 2013) decrease surface and bottom water NO₃⁻, respectively. Using the nitrate- $\Delta\text{N}_2\text{O}$ relationship to infer regression yield across wider range of oxygen levels at various marine environments (Table 1), we found that yields in Saanich Inlet were, on average, lower than those from open-ocean OMZs influenced by coastal upwelling (e.g., the Eastern tropical South Pacific) and similar to those from oxygenated oceans (e.g., Subarctic region of Pacific and Atlantic). Furthermore, a proxy of N₂O yield during nitrification is the regression slope of $\Delta\text{N}_2\text{O}$ -AOU relationship; compilation of $\Delta\text{N}_2\text{O}$ versus AOU also showed Saanich Inlet to have lower ratios of $\Delta\text{N}_2\text{O}$ versus AOU than those of open-ocean OMZs (Capelle et al., 2018; Grundle et al., 2012). Thus, we conclude that Saanich Inlet samples displayed lower N₂O yields, particularly at oxygenated conditions ($\text{DO} > 15 \text{ } \mu\text{mol L}^{-1}$) than those observed in open-ocean OMZs.

As a microbial process, N₂O production from NH₄⁺ oxidation is mediated by two distinct microbial groups, ammonium oxidizing archaea (AOA) and ammonium oxidizing bacteria (AOB). Ammonia oxidizers can be found close to the oxic-anoxic interface and can utilize nanomolar oxygen level (Bristow et al., 2016; Louca et al., 2016), which can explain the detection of nitrifying activity at 130 m in April and at 160 m in August (section 3.4). In the open ocean, the ammonia oxidizing community is dominated by AOA (Horak et al., 2018), which have been shown to demonstrate higher N₂O yields than AOB (Löscher et al., 2012), and AOA is largely responsible for N₂O production (Santoro et al., 2011). In coastal zones, the relative abundance of AOB was shown to

Table 1
Nitrification N₂O Yield Indicated by Nitrate-ΔN₂O Relationship (Referred to as Regression Yield) at Various Marine Environments

Sampling location	N ₂ O Yield (%)	O ₂ (μmol L ⁻¹)	Depth (m)	Data source
Saanich Inlet	0.039–0.051	0–226	40–150	This study and Torres-Beltrán et al. (2017)
Eastern tropical South Pacific	0.169–0.851	20–230	5–500	Ji et al. (2019)
Mauritanian upwelling	0.130–0.257	>35	0–2930	Rees et al. (2011)
Northeast Pacific	0.08	16–342	0–600	Grundle et al. (2012)
Subarctic North Atlantic	0.057 – 0.077	210–290	20–1000	Ji and Ward (2017)

Note. See regression plots in Figures S2a and S2b, respectively.

increase with increasing DO along the oxycline (Santoro et al., 2008). In contrast, N₂O yields from ammonium oxidation under suboxic conditions (DO < 15 μmol L⁻¹) in Saanich Inlet were similar to those under open ocean suboxic conditions. This similarity may indicate that the relative importance of AOA versus AOB is more comparable between coastal and open-ocean systems when suboxic conditions persist. Further analyses of the relative abundance of AOA versus AOB from oxygenated to low oxygen conditions in Saanich Inlet will help to resolve this.

5. Conclusions

Saanich Inlet is characterized by temporal and vertical oxygen gradients that modulate nitrogen speciation, N₂O distribution, and production pathways. The oxic-anoxic interface marks the boundary between net N₂O production and consumption zones, which were indicated by N₂O supersaturation and undersaturation, respectively. Above the interface, both ¹⁵N tracer incubation experiments and in situ SP values confirmed that NH₄⁺ oxidation is the dominant N₂O production pathway, probably causing N₂O supersaturation and contributing to N₂O efflux to the atmosphere. We conclude that elevated N₂O production from NH₄⁺ oxidation following the autumn renewal event contributed to the increase of water column ΔN₂O from August to October. Pronounced N₂O undersaturation within the anoxic zone indicates net N₂O consumption in the near absence of NO₃⁻ and NO₂⁻. Such a highly reducing environment could be a net N₂O sink when oxygen, NO₃⁻, and NO₂⁻ are depleted, for example, removing N₂O transported into the anoxic zone from the oxycline through advection/diffusion. As shown by incubation experiments, addition of NO₃⁻ and NO₂⁻ to anoxic samples stimulated N₂O production. Conversely, reoxygenation of anoxic samples stimulated N₂O production from NH₄⁺ oxidation. In comparison to open ocean OMZs, Saanich Inlet has high NH₄⁺ fluxes (0.6–3.5 nmol m⁻² s⁻¹) from the anoxic depths that can support nitrification and associated N₂O production. The dominance of NH₄⁺ oxidation in N₂O production in Saanich Inlet is a combination of three factors: (1) water column suboxic conditions, (2) relatively high nitrification rates (few hundred nanomole per liter per day) supported by high NH₄⁺ fluxes from the anoxic layer, and (3) depletion of NO₂⁻ and NO₃⁻ inhibiting partial denitrification to produce N₂O in suboxic and anoxic depths.

Data Availability Statement

Readers may find all the published data in PANGAEA Repository (<https://doi.pangaea.de/10.1594/PANGAEA.912191>).

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